

1 **The Middle East as a Natural Laboratory to Advance Our Understanding of Global**
2 **Hyper-Arid Drylands**

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16
17 **Abstract**

18 Contrary to the common perception of hyper-arid drylands as barren and lifeless, these regions are
19 home to some of the planet's most unique biodiversity and support over 100 million people.
20 Despite their ecological and human significance, hyper-arid drylands remain among the least
21 studied biomes in the world. In this paper, we explore how improving our understanding of hyper-
22 arid ecosystems in the Middle East can yield valuable insights applicable to other hyper-arid
23 regions. We examine how ongoing greening initiatives in the Middle East offer a unique
24 opportunity to deepen our knowledge of dryland ecology and advocate for the establishment of a
25 comprehensive research program in the region. This program would focus on ecosystem
26 functionality across spatial and temporal scales, setting the stage for a global monitoring network
27 for hyper-arid drylands. Such efforts would inform conservation strategies and climate change
28 mitigation, while also shedding light on the resilience and adaptability of hyper-arid ecosystems

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29 to environmental change. Ultimately, this monitoring would guide management practices to
30 preserve biodiversity, enhance ecosystem services, and promote sustainable development in hyper-
31 arid regions worldwide.

32

33 **Keywords:** drylands; hyper-arid; conservation; restoration; biodiversity.

34

35 Hyper-arid drylands, areas with an aridity index (precipitation/potential evapotranspiration) below
36 0.05, represent some of the most extreme environments on Earth. Despite the perception as being
37 inhospitable to life, they host a diverse set of biota and ecosystems, including rangelands that
38 provide grazing for nomadic tribes (Johnson, 1993), biocrusts that contribute to carbon
39 sequestration (Kidron et al., 2015), or coastal mangroves and salt-marshes that support fisheries
40 and modulate nutrient cycling (El-Regal and Ibrahim, 2014). Encompassing an area of around 10
41 million km², the extent of hyper-arid regions is expected to grow by the end of the century due to
42 increasing aridity driven by climate change. Current projections estimate the expansion of hyper-
43 arid land by 2050 to range from 6% under moderate scenarios to as much as 12% in the most
44 pessimistic scenarios (Huang et al. 2016). While more than 100 million people currently live in
45 hyper-arid drylands (MEA, 2005), population growth rates as high as 65% by 2100 have been
46 projected for developing countries in these regions (Huang et al., 2016), placing further strain on
47 these ecosystems.

48 Hyper-arid ecosystems remain poorly studied compared to other dryland and non-dryland
49 ecosystems (Brito et al., 2014; Šmíd et al., 2021). Research on their biodiversity, structure, and
50 function is limited, representing less than 3% of all dryland studies (Groner et al., 2023). These
51 ecosystems are not only challenging to access (Ficetola et al., 2013), but also vastly under-
52 protected, with just 6.7% of their total area designated for conservation (Lewin et al., 2024). The
53 inaccessibility of hyper-arid areas, coupled with the misconception that they are barren and devoid
54 of life, has resulted in their neglect in conservation efforts (Durant et al., 2012). Consequently,
55 there is a widespread but incorrect belief that these environments are either ecologically
56 insignificant or incapable of further degradation (Martínez-Valderrama et al., 2020). Contrary to
57 this view, hyper-arid drylands are rich in biodiversity. For example, the Algerian Sahara alone is
58 home to at least 1,200 plant species (Ozenda, 2004). Due to the unique adaptations of organisms
59 in these extreme environments, hyper-arid ecosystems offer valuable insights into how dryland
60 systems might respond to future climate change. They serve as natural laboratories for studying
61 the impacts of, and adaptations to, climatic change that could affect other dryland regions (Groner
62 et al., 2023; Grünzweig et al., 2022). Furthermore, as non-dryland areas face increasing water
63 scarcity, mechanisms governing ecosystem functioning in drylands are expected to become
64 relevant in these regions (Allan et al., 2020). Many of these changes are anticipated in densely
65 populated regions, particularly in the subtropics and mid-latitudes, with significant implications
66 for food production and societal well-being (Grünzweig et al., 2022). Beyond ecological insights,
67 studying the adaptations of organisms in hyper-arid drylands holds great promise for
68 biotechnological and biodiversity applications (Bull and Asenjo, 2013).

69 The Middle East accounts for over 30% of the world's hyper-arid drylands. This region
70 hosts diverse biomes that have developed unique eco-evolutionary adaptations over thousands of
71 years of biotic and abiotic interactions. They support more than 8,000 unique species of vascular
72 plants (Hegazy and Doust, 2016) and encompass a diverse range of ecosystems present in other
73 regions, albeit under more favorable conditions. These ecosystems span from mangroves along the
74 coastal fringes of the Red Sea to grasslands and shrublands extending across Turkey and Iraq (**Box**
75 **1**). Despite the geographic and historical interest the Middle East has generated, much of the
76 research undertaken in this region has predominantly focused on the description of the flora in
77 individual countries, such as Iran (e.g., Rechinger, 1963-2005), Israel and Palestine (e.g., Zohary,
78 1962; Danin, 2004), Lybia (e.g., Jafri and El-Gadi, 1977-1993), Oman (e.g., Ghazanfar, 1992,
79 1998), Saudi Arabia (e.g., Migahid, 1978; Mandaville, 2013), Turkey (e.g., Davis, 1984; Davis et
80 al., 1988), or Yemen (e.g., Kilian et al., 2002; Brown and Mies, 2012). Other studies described the
81 vegetation of the Middle East from geobotanical and phytogeographical perspectives (e.g., Zohary,
82 1971, 1973). Similarly, the fauna of the Middle East has drawn significant interest due to the
83 extreme environmental conditions these species endure, with several biodiversity hotspots in the
84 region. For example, the Arabian Peninsula hosts a high number of endemic vertebrate species,
85 21.6% of which are unique to this region (Mallon, 2011). Additionally, the Middle East serves as
86 an essential stopover for migratory bird species along major migratory routes that connect Africa,
87 Asia and Europe (Schekler et al., 2022). Countries like Israel have been extensively studied for
88 their key role in bird migration routes for decades already (e.g., Leshem and Yom-Tov, 1996).
89 However, except for Hegazy and Doust (2016), the life stories of many Middle Eastern species
90 have not been comprehensively investigated and described while concurrently considering this
91 region's geography, plant evolution and ecology. Moreover, these studies have yet to integrate the
92 complex interactions between human societies and ecosystems, particularly in the face of the
93 additional pressures imposed by climate change.

94 Here, we elaborate on how research on the biodiversity and ecology of Middle East hyper-
95 arid drylands can advance our understanding of dryland ecosystems globally, while also
96 contributing to the success of ongoing Saudi and Middle East Green Initiatives
97 (<https://www.greeninitiatives.gov.sa/>). With an initial investment of more than USD 180 billion,
98 these green initiatives aim to restore degraded marine and terrestrial environments, enhance
99 biodiversity and mitigate the impacts of climate change throughout the Middle East. We argue that
100 if these initiatives are successfully developed and implemented, they might serve as the foundation
101 for further experimental and theoretical studies on the impacts of extreme climates on dryland
102 ecosystems globally. Furthermore, the Saudi and Middle East Green Initiatives could establish the
103 base for applied solutions aimed at preserving and/or rehabilitating the biodiversity and ecosystem
104 services of global drylands, mitigating climate change and addressing land degradation and
105 desertification.

106

107 **Ongoing greening initiatives in the Middle East: An untapped potential to enhance our**
108 **understanding of hyper-arid ecosystems**

109

110 To sustain its unique biodiversity into the future, it is crucial to promote the resilience and health
111 of hyper-arid ecosystems, particularly given the compound pressures of anthropogenic influence
112 and climate change. These are key objectives of the Saudi and Middle East Green Initiatives, which
113 aim to protect up to 30% of Saudi Arabia's land and sea territories and plant up to 10 and 40 billion
114 trees within the Kingdom and across the Middle East, respectively
115 (<https://www.greeninitiatives.gov.sa/about-sgi/> and [https://www.greeninitiatives.gov.sa/about-](https://www.greeninitiatives.gov.sa/about-mgi/)
116 [mgi/](https://www.greeninitiatives.gov.sa/about-mgi/), respectively). Other actions supported by these initiatives include the increase of renewable
117 energy capacity – which has already risen by 300% in Saudi Arabia-, restoring degraded lands –
118 94,000 hectares have been rehabilitated across Saudi Arabia at the moment – and rewilding
119 endangered species that play a key role in the ecological balance of these ecosystems. It is also
120 expected that the Saudi Green Initiative will play a significant role in achieving the recent
121 commitment of Saudi Arabia to reach net zero emissions by 2060, with the Middle East Green
122 Initiative aiding broader regional objectives towards carbon neutrality. Moreover, the Saudi Green
123 Initiative's ambition to protect at least 30% of Saudi Arabia's territories by 2030 is in harmony
124 with the global "30x30" target adopted under the Kunming-Montreal Global Biodiversity
125 Framework of the Convention on Biological Diversity (CBD, 2022). Although this is a challenging
126 objective, 18.1% and 6.49% of Saudi Arabia's terrestrial and marine areas are already protected.

127 Both greening initiatives will protect some of the region's iconic terrestrial fauna, which
128 are classified at varying levels of threat, ranging from vulnerable to critically endangered,
129 according to the International Union for Conservation of Nature (IUCN) Red List of Threatened
130 Species. Additionally, they will protect mangroves, coral reefs and salt marshes, which have co-
131 evolved in this region to create some of the most resilient marine ecosystems globally (McCabe et
132 al., 2023). Some of the Saudi Green Initiative activities include the creation of national reserves,
133 such as the King Salman bin Abdulaziz Royal Reserve, located in the north of the Arabian
134 Peninsula. Covering approximately 130,000 km², this reserve hosts vulnerable species of
135 mammals (e.g., *Capra nubiana*, *Canis lupus arabs*) and birds (e.g., *Torgos tracheliotos*, *Falco*
136 *cherrug*). Additionally, urban areas are targeted by these initiatives. Cities like Riyadh and Makkah
137 in Saudi Arabia are seeing an increase in the number of trees planted and the creation of new green
138 areas, enhancing human well-being and biodiversity (Gaston, 2010; Cox et al., 2017).

139 The Saudi and Middle East Green Initiatives should also learn from past actions and seek
140 not only to ecologically transform broad landscapes but also to shape societies and economies. For
141 example, the Great Green Wall for the Sahara and the Sahel Initiative (GGWSS), which emerged
142 in 2007, involves over 20 countries bordering the Sahara to establish plantations on 100 million
143 ha from Eritrea's Red Sea coast to Senegal's Atlantic coast (Sileshi et al., 2023). The GGWSS built
144 upon earlier initiatives aimed at combating desertification in the Sahel region's countries (Mbow,
145 2017). One such initiative was Algeria's Green Dam Initiative, started in 1972, which aimed to
146 establish a three million ha band of plantations to halt the northward advance of the Sahara Desert
147 (Benhizia et al., 2021). Other projects, such as the Acacia operation project and the Support for
148 the rehabilitation and extension of the Nouakchott green belt in Mauritania, engaged local
149 communities and national authorities in restoring inland and coastland ecosystems (Berte, 2010).
150 Projects in the Sahel region have shown that where policies and incentives are favorable, farmers
151 actively promote the natural regeneration of trees, resulting in vast areas now being covered by
152 trees (e.g., Haglund et al., 2011). A participatory approach involves extensive community
153 engagement and enhances accountability and stewardship in land-restoration efforts. Initially, a

154 centralized approach, heavily reliant on forest department control and substantial investment in
155 equipment, marginalized local communities. Recognizing community ownership has enabled
156 Sahelian countries to mitigate conflicts between development and environmental goals (Kumar,
157 2003). However, land privatization in the Sahel often fails due to diverse landscape uses and
158 stakeholder needs (Schoneveld, 2017). These failures underscore the necessity for stakeholder-
159 supported, site-specific solutions that enable ongoing improvement across countries and
160 implementation sites. Learning from experiences in the Sahel region, local actions that can be
161 scaled up with positive results include the zoning of grazing areas, ensuring water availability for
162 livestock, and promoting fodder trees (Mbow, 2017).

163 In Asia, the Great Green Wall of China (GGWC), initiated by the Chinese government in
164 1978, aims to combat desertification and reduce the eolian transport of dust from the Gobi Desert
165 (Parungo et al., 1994). Scheduled for completion in 2070 (Lu et al., 2018), this project builds on
166 China's experience with shelterbelt programs (Qi and Dauvergne, 2022). While the GGWC has
167 yielded benefits, such as reduced dust movement and increased vegetation, during its first stages,
168 many of the dryland areas targeted for afforestation were found to be better suited for grasslands
169 and steppes than woodlands or forests (Cao et al., 2010; Mátyás et al., 2013), often leading to
170 significant water pressures on water resources (Li et al., 2021a). Not only tree survival rates were
171 low but also irrigation was necessary in drier areas within many of these projects (e.g., Wang et
172 al., 2020). Nevertheless, subsequent research has demonstrated the benefits of shelterbelts in
173 drylands for reducing net erosion (Su et al., 2021) and improving crop productivity (Zheng et al.,
174 2016). Additionally, studies on biocrusts in China's drylands have shown that breeding them can
175 effectively control land degradation (Li et al., 2021b) by reducing dust emissions and increasing
176 soil nutrient content (Li et al., 2010; He et al., 2019). Because of these experiences, new strategies
177 in China now focus on science-based activities, encouraging natural regeneration, creating multi-
178 species plantations, matching species to local conditions and emphasizing water conservation
179 (Turner et al., 2023).

180 Over the past four decades, Australia has also made significant advancements in restoring
181 its drylands through sustained efforts and community involvement (Campbell et al., 2017).
182 Initiatives in Australia learned from small-scale efforts and led to a shift in policies towards large-
183 scale activities, biodiversity conservation, water quality improvement and greenhouse gas
184 mitigation. Successful restoration programs underscored community capacity and commitment,
185 yet it was also recognized that community efforts alone were insufficient for sustainable resource
186 management on a landscape or continental scale without technically and economically viable land
187 use and farming systems. These lessons are particularly important in drylands, where synergistic
188 interactions such as grazing intensification, drought, climate change, reduced fire frequency, and
189 changes in atmospheric chemistry or small animal populations can collectively overwhelm the
190 effects of individual factors (Fu et al., 2021a).

191

192 **Restoration in the Middle East cannot be based only on planting trees in the desert**

193 Ambitious tree-planting objectives are not a new concept, even in drylands (Bond et al., 2019).
194 Unfortunately, many previous dryland afforestation efforts have often delivered tree monocultures,

195 which risks reducing sustainable development by negatively affecting ecosystem functioning (Yao
196 et al., 2021). Apart from avoiding planting regimes that are incompatible with the landscape, the
197 inherent constraints of water availability in drylands, and the increased pressures that large-scale
198 tree planting places on these, are critical considerations when designing greening and restoration
199 efforts (Schwärzel et al., 2020). Although intrinsically appealing from a policy perspective (i.e.,
200 planting trees is a socially recognizable and acceptable climate action), excessive focus on
201 afforestation using trees can miss opportunities for broader and longer-term benefits. For instance,
202 mono-specific tree plantations may achieve a narrow accounting-based objective (in terms of trees
203 planted or carbon captured) but they can reduce ecosystem diversity (e.g., Maestre and Cortina,
204 2004), jeopardize water resources for humans and ecosystems (e.g., Feng et al., 2016), and amplify
205 the risk of future carbon loss following any ecosystem disturbance (e.g., forest fires and pests;
206 Anderegg et al., 2020). In other parts of the world, regions deemed degraded have been mistakenly
207 considered as potential areas for afforestation, simply by failing to carefully assess their suitability
208 for tree planting (e.g., soil health, environmental gradients). Such areas have included grasslands
209 and shrublands (Veldman et al. 2019), which represent two of the more common environments
210 found in the Middle East (Box 1; Hegazy and Houst, 2016).

211 Recognizing the limitations and unintended consequences of prior afforestation strategies
212 underscores the importance of adopting a more nuanced approach to ecosystem restoration,
213 particularly in hyper-arid regions. Increased biodiversity is considered an indicator of healthier
214 and more resilient ecosystems, allowing faster recovery from disturbance and providing ecosystem
215 services that contribute to more sustainable and stable human development (Jactel et al., 2017).
216 Thus, restoration and conservation efforts should act in concert to increase biodiversity, thereby
217 bolstering the resilience of all naturally occurring ecosystems. This holistic view is crucial if the
218 goal is to restore the multifaceted ecosystems of hyper-arid lands, considering the variety of
219 services they provide (Box 1). For example, biocrusts are key players in dryland development and
220 function that increase soil carbon and nutrient contents, impact multiple components of the
221 hydrological cycle, and reduce soil erosion and dust emissions (Eldridge et al., 2020; Rodríguez-
222 Caballero et al., 2022), benefitting both the environment and human societies. Therefore, the
223 development of a biocrust research program is urgently needed to understand their ecology,
224 distribution and potential to restore degraded habitats and mitigate climate change in the Middle
225 East.

226 Restoration and greening initiatives in the Middle East should focus not only on what is
227 visible above ground but also on soils. Over 32% of the world's soil organic pool is stored in
228 drylands worldwide (Plaza et al., 2018a), with significant loss of carbon occurring in major
229 cropland and grazing areas (Sanderman et al., 2017). However, although soils' potential to mitigate
230 climate change has been long recognized (Bossio et al., 2020), their role in dryland restoration and
231 mitigation efforts remains underexplored. Soil organic carbon can act as a stable carbon sink,
232 showing resilience to land-use changes and disturbances, unlike above-ground biomass. Carbon-
233 rich soils also enhance water and nutrient retention, enhancing ecosystem resilience to disturbances
234 like droughts (e.g., Izumi and Wagai, 2019). However, regional evaluations of soil organic carbon
235 in drylands remain limited, with existing studies often producing inconsistent results (Fu et al.,
236 2021b). Furthermore, understanding the impact of land-use changes on regional soil carbon is
237 hindered by insufficient data quality, poor representativeness, and a lack of historical land-use
238 information (Hendriks et al., 2016). Comprehensive assessment of soil carbon stocks requires

239 robust sampling methods that can scale site-specific data to broader regional levels (Ciais et al.,
240 2011), an ongoing challenge in terrestrial carbon studies (Zhang and Hartemink, 2017). As such, a
241 regional dataset combining soil organic carbon, land-use, and soil properties for the Middle East
242 would enhance our understanding of how climate influences physical processes in global drylands.
243 For instance, increasing aridity is known to reduce soil carbon and nitrogen levels (Delgado-
244 Baquerizo et al., 2013) and disrupt nutrient balance in dryland soils (Maestre et al., 2016). Carbon
245 accumulation in soils is influenced by factors such as parent material, topography, microclimatic
246 conditions, and species diversity (Ramesh et al., 2019), while human activities can accelerate
247 carbon emissions (Schlesinger, 2000; Lal, 2004a). Although improved management strategies
248 (e.g., grazing regimes, organic amendments, cover crops, crop rotation and conservation tillage)
249 can enhance carbon stocks in dryland soils (Plaza et al., 2018b; Lal, 2004b; Lal, 2018), they can
250 be less effective in these environments due to their coarser texture and lower clay content, which
251 protects organic matter from decomposition (Six et al., 2002; Lehmann and Kleber, 2015). It is,
252 therefore, crucial to evaluate the interactions between biotic, abiotic and human factors to
253 understand soil C dynamics in the Middle East.

254

255 **Concluding remarks**

256 Hyper-arid lands have been largely missing from existing large-scale global dryland field surveys
257 (Maestre et al., 2012, 2022b). The Saudi and Middle East Green Initiatives provide a unique
258 opportunity to gain insights into the processes that govern the structure, functioning, and responses
259 to climate change of hyper-arid drylands. Knowledge gaps that need to be addressed include
260 understanding: i) the drivers for the unexpected high functional diversity in dryland plants (Gross
261 et al., 2024); ii) how plants will adapt to water scarcity and respond to increased inter-annual
262 precipitation variability (Garcia-Pichel and Sala, 2022); iii) developing a region-wide
263 understanding of the distribution, characteristics, and functioning of biocrusts (e.g., Abed et al.,
264 2019), and iv) the mechanisms, both physiological and genetic, behind the ability of soil
265 microorganisms to endure extreme conditions (Makhalanyane et al., 2015). Further, many remote
266 sensing-derived products ignore hyper-arid drylands based on the assumption that vegetation is
267 largely absent. As a result, hyper-arid drylands are often excluded from remote sensing products
268 typically used in global studies and vegetation estimates (e.g., Harris et al., 2021; Sabatini et al.,
269 2022). This is problematic, as vegetation (and trees in particular) is more abundant in hyper-arid
270 areas than initially thought (Brandt et al., 2020; Reiner et al., 2023). More generally, international
271 networks evaluating ecosystem carbon, water, and energy fluxes, such as FLUXNET (Balocchi
272 et al., 2001), lack sites in hyper-arid environments, despite these representing around 8% of global
273 land surface (Právělie et al., 2019). Developing an augmented flux network that includes sites in
274 the Middle East would provide invaluable information on hyper-arid drylands, and contribute to
275 fill existing gaps in flux databases that preclude obtaining more precise carbon cycling and climate
276 change impact estimates.

277 Many of the actions discussed here will occur in complex and unpredictable contexts,
278 where human realities should be considered alongside ecological and biophysical factors. Drylands
279 exhibit sensitivity to changes in structure–function relationships due to extreme climate conditions
280 (Reynolds et al., 2007; D’Odorico and Bhattachan, 2012), and human interventions can alter the

281 resilience and stability of these systems (Robinson et al., 2015). Interactions between natural and
282 human-induced processes affect dryland dynamics at specific scales (Fu et al., 2021a), varying by
283 social, cultural, and economic context (Stringer et al., 2017). Therefore, understanding the
284 complex and adaptive nature of drylands in the Middle East should involve dynamic interactions
285 between its ecosystems and human societies (Folke et al., 2016; Schütler et al., 2019), requiring
286 interdisciplinary efforts (Bautista et al., 2017; Schütler et al., 2019).

287 Research in the Middle East should focus on the interplay between ecosystem services and
288 human well-being to optimize services that enhance drylands' health and human well-being in the
289 long term (Fu et al., 2021b). Identifying local limiting factors and their impacts can improve
290 knowledge of ecosystem functioning and livelihoods through sustainable development (Turner et
291 al., 2003; Reed et al., 2015). An interdisciplinary approach, which evaluates ecological and social
292 perspectives together, will allow for assessing ecological dynamics and their driving forces in the
293 Middle East. It will not only enable an understanding of the macroscopic differences among
294 various dryland systems in this region but also help identify management or policy responses likely
295 to deliver successful outcomes in different types of drylands (Fu et al., 2021b).

296 Developing a comprehensive research program on ecosystem structure and functioning
297 across multiple spatio-temporal scales in the Middle East is a critical step to provide the scientific
298 underpinning needed for the success of ongoing green initiatives and climate change and
299 desertification mitigation actions in this region. The creation of a Middle East collaborative
300 network of researchers, practitioners, and decision-makers, and the set-up of standardized regional
301 surveys using standardized protocols following models successfully implemented in other large-
302 scale and global surveys (e.g., Maestre & Eisenhauer 2019; Maestre et al., 2022a), would be a
303 fundamental step forward towards achieving this aim. Doing so would not only be key to creating
304 the basis for long-term monitoring of ecosystem changes in the region, but would provide us with
305 invaluable insights to advance our understanding of hyper-arid drylands and to better comprehend
306 and react to the increasingly drier conditions being experienced and forecasted across the globe.

307

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311

312 **Author Contribution Statement**

313 J. B.-S. conceived the idea and wrote the initial draft; all authors contributed to the review and the
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315

316 **Conflict of Interest Statement**

317 The authors declare that they have no conflict of interest.

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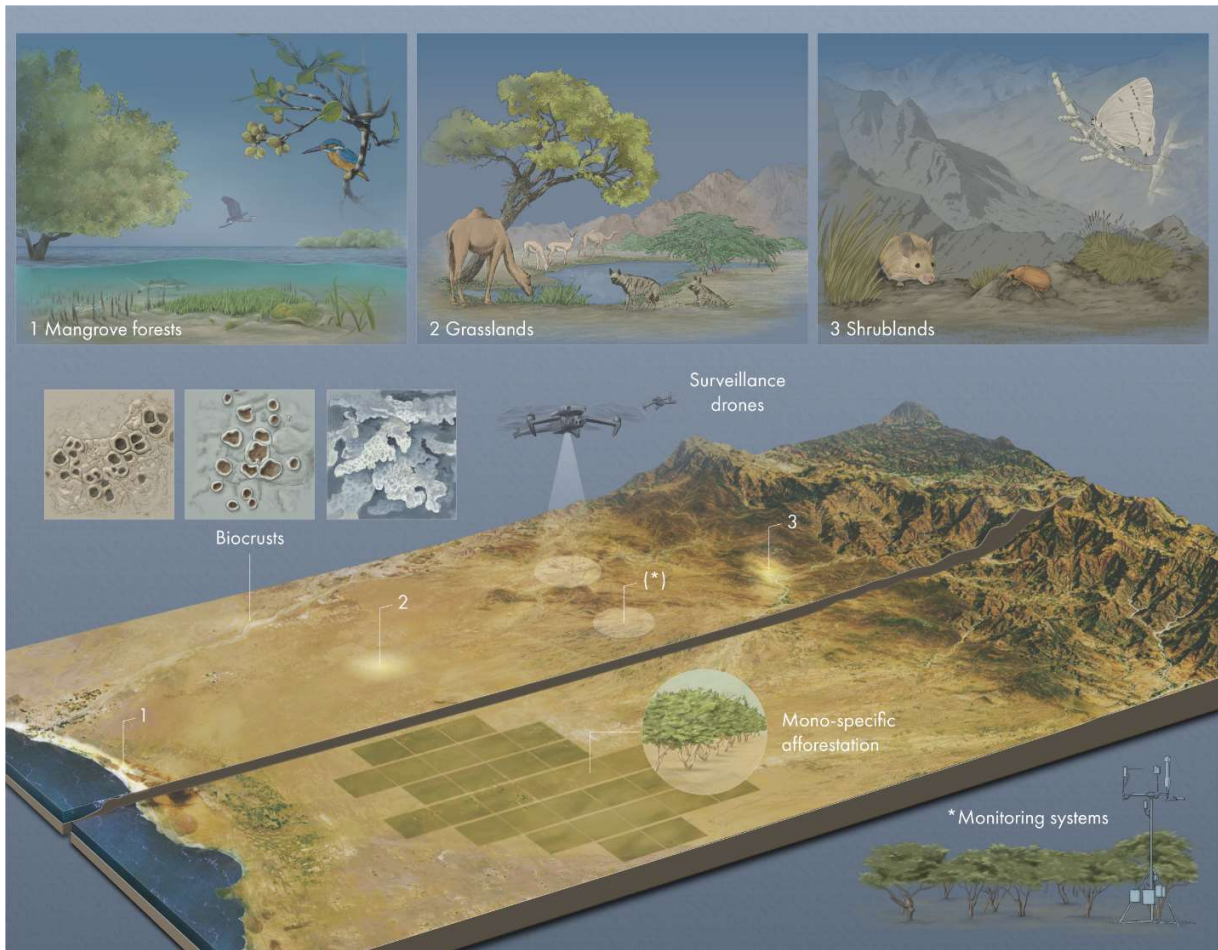
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657 Box 1. Hyper-arid drylands in the Middle East are much more than barren landscapes.

658 The Middle East is home to diverse ecosystems that, while also found in other regions under more
659 favorable conditions, can thrive in some of the driest environments on Earth. Gaining a deeper
660 understanding of these ecosystems offers valuable insights into their functioning, restoration
661 potential, and relevance for addressing climate change, land degradation, and desertification. For
662 instance, mangroves (1) along the coasts of the Red Sea and Arabian Sea (Almahasheer et al.,
663 2018; Blanco-Sacristán et al., 2022) are a key vegetation type in the Middle East. These ecosystems
664 provide nursery grounds for marine life, support local communities through commercial species,
665 and protect coastlines from erosion. However, mangroves in this region endure extreme saline
666 stress due to limited freshwater inputs and increasing groundwater extraction, on which they
667 heavily rely (Adame et al., 2021). Additionally, these mangroves face significant human pressures
668 (Almahasheer et al., 2016). Understanding how mangroves survive in such arid conditions offers
669 a unique opportunity to predict how global mangrove ecosystems might respond to climate change,
670 including the effects of human activities, sea-level rise, and microclimatic shifts (Osland et al.,
671 2016). Similarly, the grasslands (2) of the Middle East, such as those in Iran's Taftan mountains
672 (Burrascano et al., 2018) and the southwestern Arabian Peninsula (Ghazanfar and Fisher, 1998),
673 provide a valuable opportunity to study the interactions between abiotic and biotic factors across
674 altitudinal and latitudinal gradients. As global aridity increases, understanding the dynamics of
675 these grasslands—ranging from Mediterranean grasslands to semi-arid steppes—can offer crucial
676 insights for improving grassland health in other dryland regions. This is particularly important
677 given the extreme climatic conditions in which these grasslands exist, which mirror those in many
678 other arid and semi-arid ecosystems globally, such as the grasslands in the Namib Desert (Evans
679 et al., 2020; Logan et al., 2021) and Australia (Keast, 2013). By studying how these Middle Eastern
680 grasslands thrive, researchers can gain a deeper understanding of the resilience and adaptive
681 strategies of grassland ecosystems, crucial for managing the effects of climate change on
682 grasslands worldwide. Shrublands (3), which dominate much of the Middle East—such as the
683 eastern Arabian Peninsula, parts of Jordan (e.g., Jebel Ajloun), and northern Israel (Upper
684 Galilee)—also play a critical ecological role. They provide habitats for insects and small rodents
685 and host biocrusts—communities of photo- and heterotrophic organisms living on the soil surface
686 in large, unvegetated drylands. Biocrusts are essential for maintaining dryland ecosystem health
687 by influencing soil respiration, nutrient cycling, and runoff dynamics. While biocrusts have been
688 extensively studied in regions like the Negev Desert and the Arava Valley in Israel (e.g., Galun
689 and Garty, 2003; Kidron and Tal, 2012), research across other Middle Eastern countries is limited.
690 Studies from countries like Iran (Bashtian et al., 2019), Iraq (Hamdi et al., 1978), Jordan (El-Oqlah
691 et al., 1986), Oman (Abed et al., 2013), and Saudi Arabia (Alotaibi et al., 2020) suggest that
692 biocrust composition is relatively uniform across the region (Galun and Garty, 2003), but more
693 research is needed to fully understand their distribution and composition in the Middle East. With
694 its long history of land use and anthropogenic impacts under climate change (Kaniewski et al.,
695 2012), the Middle East offers valuable insights into how human activities shape biocrust
696 communities under extreme environmental conditions. Studying the interactions between biocrusts
697 and human-induced changes—such as grazing, agriculture, and urbanization—can inform
698 strategies for managing and mitigating these impacts, both regionally and globally. Leveraging
699 remote sensing technologies (e.g., satellites, drones, and eddy-covariance towers) alongside in-situ
700 data collection could enhance ecosystem surveys, providing timely insights into their functioning.
701 This data could also support the establishment of new monitoring networks, such as eddy

702 covariance flux networks, which remain underrepresented in hyper-arid drylands worldwide
703 (Smith et al., 2019).

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